



cambridge.org/enc

## Research Paper

**Cite this article:** Plieninger T et al. (2026) Local peoples' values and disvalues in and around an Indian protected area undergoing urbanization. *Environmental Conservation* 53: 40–48. doi: [10.1017/S0376892925100234](https://doi.org/10.1017/S0376892925100234)

Received: 20 June 2025

Revised: 11 October 2025

Accepted: 15 October 2025

First published online: 10 November 2025

### Keywords:









Conservation conflicts; free-listing; India; protected areas; Public Participation Geographic Information Systems (PPGIS); urbanization

### Corresponding author:

Tobias Plieninger;

Email: [plieninger@uni-goettingen.de](mailto:plieninger@uni-goettingen.de)

# Local peoples' values and disvalues in and around an Indian protected area undergoing urbanization

Tobias Plieninger<sup>1</sup> , Pramila Thapa<sup>1,2</sup> , Nora Fagerholm<sup>3</sup> , Sukanya Basu<sup>1</sup> , Dhanya Bhaskar<sup>4</sup> , Harini Nagendra<sup>5</sup> , Christopher M Raymond<sup>6,7,8</sup>  and Mario Torralba<sup>1,9</sup> 

<sup>1</sup>Social-Ecological Interactions in Agricultural Systems, University of Göttingen and University of Kassel, Witzenhausen, Germany; <sup>2</sup>Social-Ecological Systems Institute, University of Lüneburg, Lüneburg, Germany; <sup>3</sup>Department of Geography and Geology, University of Turku, Turku, Finland; <sup>4</sup>Faculty of Ecosystem and Environment Management, Indian Institute of Forest Management, Bhopal, India; <sup>5</sup>School of Climate Change and Sustainability, Azim Premji University, Bengaluru, India; <sup>6</sup>Helsinki Institute for Sustainability Science, University of Helsinki, Helsinki, Finland; <sup>7</sup>Ecosystems and Environment Program, Faculty of Biological and Environmental Sciences, University of Helsinki, Helsinki, Finland; <sup>8</sup>Department of Economics and Management, Faculty of Agriculture and Forestry, University of Helsinki, Helsinki, Finland and <sup>9</sup>IVM Institute for Environmental Studies, VU University Amsterdam, Amsterdam, The Netherlands

## Summary

Urbanization has become a key pressure on many of the world's protected areas. This study investigates how local communities perceive landscape values and disvalues in and around Bannerghatta National Park (near Bengaluru, India), which is experiencing high rates of urban development at its peripheries. Using combined free-listing and Public Participation Geographic Information Systems (PPGIS) mapping, we surveyed 489 residents from 12 villages to elicit both landscape values and disvalues. Respondents mapped values such as biodiversity, fertile land and clean air, while disvalues focused on human–wildlife conflicts. Despite persistent conflicts and urbanization pressures, residents valued the National Park for its multiple landscape values. Both values and disvalues were concentrated around village areas. We find that socio-demographic factors – especially caste, land ownership and work in agriculture – significantly influenced perceptions. Specifically, marginalized caste members and landless individuals reported more disvalues, while landowners and farmers noted more values. Our study emphasizes the need to consider both landscape values and disvalues for balanced decision-making in protected areas. It also highlights the potential of free-listing to identify the well-being aspects that matter most for people, which points to the importance of agricultural uses in and around protected areas undergoing urbanization.

## Introduction

Comprising c. 46.4 million km<sup>2</sup> in 2020, protected areas (PAs) have become a central pillar of global conservation (Gurney et al. 2023). However, existing PAs face multiple challenges, particularly regarding social-ecological effectiveness, equitable management, ecological representativeness and connectivity (Gaston et al. 2008). Many large PAs are adversely affected by negative developments on surrounding lands, and PA management is often weak and insufficiently orientated towards biodiversity conservation needs (Palomo et al. 2014).

Although PAs are often situated in remote marginal areas (such as deserts, marshlands or mountains; Joppa & Pfaff 2009), urbanization has progressively become a key pressure (González-García et al. 2022). It has been estimated that urban land near PAs will increase from 450 000 km<sup>2</sup> in 2000 to 1 440 444 km<sup>2</sup> in 2030 (Güneralp & Seto 2013). Asia and Africa are the continents where urbanization rates are currently the highest (Nagendra et al. 2018). India is experiencing a high magnitude of urban expansion near PAs. Current and future urban pressure on PAs and their surroundings implies localized but significant degradation of biodiversity at the levels of ecoregions and individual species of conservation concern (McDonald et al. 2019).

People living in or near PAs often depend on local natural resources and are deprived of protein, timber and income sources when their livelihood activities are restricted (Loos 2021), ultimately leading to local opposition to conservation (Holmes 2013). At the same time, people living in areas surrounding PAs are frequently exposed to ecosystem disservices (defined as 'ecosystem generated functions, processes, and attributes that result in perceived or actual negative impacts on human well-being'; Shackleton et al. 2016) from PAs – for instance, crop raiding or attacks on livestock by wild mammals. While disservices focus on the instrumental

© The Author(s), 2025. Published by Cambridge University Press on behalf of Foundation for Environmental Conservation. This is an Open Access article, distributed under the terms of the Creative Commons Attribution licence (<https://creativecommons.org/licenses/by/4.0/>), which permits unrestricted re-use, distribution and reproduction, provided the original article is properly cited.

values of nature, the broader concept of ‘disvalues’ has been developed to recognize the potential for negative values across instrumental, intrinsic and relational values (Lliso et al. 2022).

Land-use conflicts that result from the perceived or actual disvalues of PAs are accelerated by urbanization impacts, such as human population growth, loss of rural identities and increased mobility. On the other hand, PAs close to urban areas can offer important contributions to people’s well-being – for example, by providing fresh water, recreational spaces and access to nature (Brill et al. 2017). In some cases, urbanization may even present opportunities for biodiversity – for example, by reducing the intensity of fuelwood collection or by offering nodes for ecosystem recovery (Nagendra et al. 2013).

Participatory approaches, such as participatory mapping, can play a critical role in helping identify the diverse values of nature in relation to PA management. A participatory mapping approach, Public Participation Geographic Information Systems (PPGIS), has increasingly been used to inform land-use planning and management of both urban and rural environments and also to help us understand the multiple dimensions of human–nature relationships (Brown & Fagerholm 2015). In the context of PAs, the mapping of visitors’ perceptions regarding nature experiences, perceived ecosystem services, tourism impacts, management preferences or infrastructural needs has often been at the centre of PPGIS studies (e.g., Solé et al. 2025). Social and cultural landscape values, land-use preferences and acceptance/rejection of management practices have also been mapped to identify potential for land-use conflicts (Engen et al. 2018). PPGIS has been used in case studies to elicit landscape uses such as harvesting of forest products or recreational activities (Torralba et al. 2025) as well as similar and dissimilar values of nature among different stakeholders (Loc et al. 2021). Ecosystem services (both social perceptions and biophysical flows) inside and outside of PAs have also been mapped at landscape levels (Palomo et al. 2013), and in some cases PPGIS has been part of stakeholder participation activities (Ioki et al. 2019).

Bannerghatta National Park (BNP) in southern India is an iconic PA that illustrates many conservation challenges. It forms a part of a critical ecological link between the Eastern and Western Ghats (a global biodiversity hotspot) and serves as a migratory corridor for an endangered flagship species, the Asian elephant (Venkataramana et al. 2015). Parts of BNP and its surroundings are inhabited by local people. Several villages inside and around BNP harbour traditionally forest-dependent communities (Gopalakrishna et al. 2010) who use the landscape for collecting wild fruits, meat, manure, firewood, lac (a resinous substance produced by insects) and honey (Varma et al. 2009). BNP is also inhabited by several tribal and marginalized caste communities (Scheduled Caste, Scheduled Tribe). Furthermore, BNP is strongly affected by the expansion of the megacity Bengaluru. Urban development has led to fragmentation of the ecological network and increased incidences of human–wildlife conflicts, especially with elephants at the fringes of BNP. With all these drivers affecting it, BNP offers important insights that are of relevance for other PAs facing urbanization in the Global South. While some studies have investigated human–elephant conflicts (e.g., Venkataramana et al. 2015) and land-cover changes (e.g., Adhikari et al. 2015) in the study area, little is so far known about the broader societal values and disvalues of BNP.

The overall aim of this study is to explore, quantify and map the interplay of landscape values and disvalues as perceived by residents living along the boundary of BNP through a free-listing

and PPGIS survey. Our specific objectives are to: (1) explore the places in the landscape that local residents valued positively and negatively (i.e., disvalued); (2) compare the frequency of landscape values and disvalues mapped inside and outside BNP; (3) identify the spatial patterns of intensity, richness and diversity of landscape values and disvalues; and (4) reveal the socio-demographic characteristics influencing landscape values and disvalues.

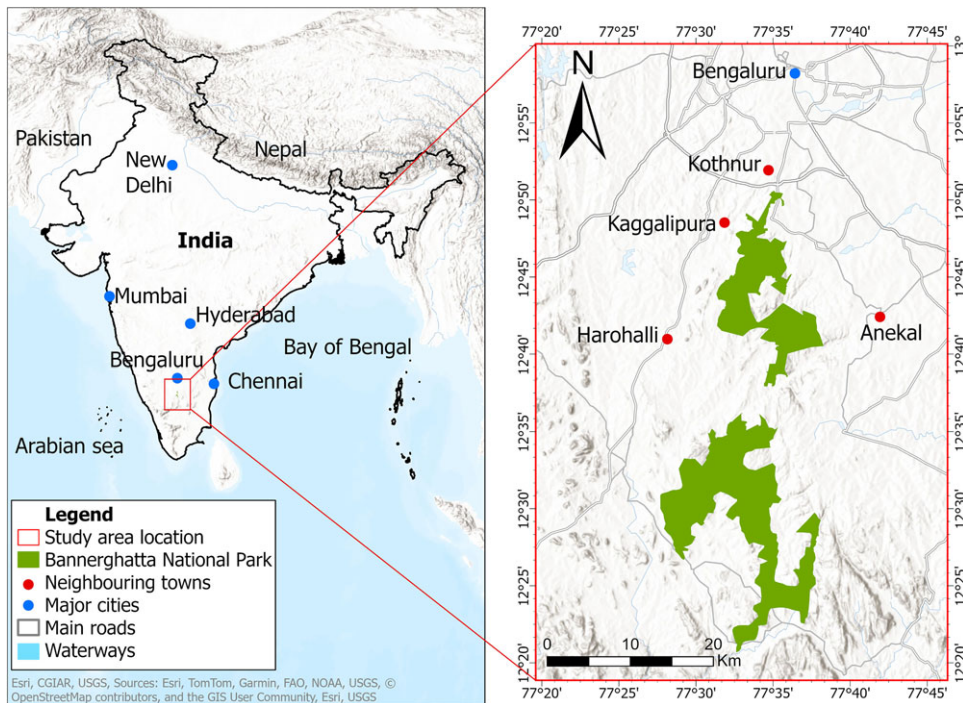
Our study complements the aforementioned participatory mapping exercises around PAs by expanding PPGIS to a joint assessment of perceived landscape values and disvalues. Stronger consideration of the positive and negative values of nature is considered a prerequisite to socially inclusive biodiversity conservation but has rarely been achieved to date (Lliso et al. 2022). Integrating landscape values and disvalues could be particularly helpful for inventorying human–nature-related issues in a specific conservation context (Oostvogels et al. 2024) and for understanding broader human–nature connections and disconnections more comprehensively (Beery et al. 2023). Failure to adequately integrate disvalues may, especially in a PA setting, lead to the alienation of people who have experienced harms from nature and/or from PA management (Luque-Lora 2024). A second contribution of our study to the idea of socially inclusive conservation (De Pourcq et al. 2019, Raymond et al. 2022) is the combination of a standardized PPGIS survey with an open-ended, qualitative elicitation of landscape values and disvalues. The open-ended nature of free-listing allows for the exploration of underlying, broader narratives regarding the values and disvalues of the landscape (c.f. Wartmann et al. 2018), while PPGIS focuses on the spatial patterns of a specific set of key values and disvalues. This dual approach enables a comprehensive analysis of landscape values and conflicts.

## Methods

### Study area

BNP is located at the southern periphery of Bengaluru, India’s fastest-growing metropolis, with a population of c. 10 million (Fig. 1). BNP is Bengaluru’s largest natural landscape, and it harbours dry deciduous and scrub forests spread over rocky and uneven landscapes. BNP was declared a national park in 1974 under the Wildlife Protection Act of 1972. The initial park area of 103 km<sup>2</sup> was expanded in 2011 to a total area of 260 km<sup>2</sup> under protection. Presently, BNP comprises 12 reserve forests spread over the districts of Bengaluru Urban, Bengaluru Rural and Ramanagara, as well as four wildlife ranges or administrative zones: Bannerghatta, Anekal, Harohalli and Kodihalli. BNP has a highly irregular shape, with a length of c. 59 km and a width varying from 0.3 km to 13.8 km. An ‘eco-sensitive zone’ was established around BNP in 2018 to cushion the National Park against developmental pressures, under which commercial activities such as mining, setting up of polluting industries or discharge of harmful effluents are prohibited, while small-scale establishments such as hotels and resorts, small industries, agriculture or road-widening projects are regulated.

Major land-cover changes have taken place around BNP, including conversion of native forest into land dedicated to agriculture, horticulture, industrial and suburban development, mining, grazing and illegal logging (Adhikari et al. 2015). The conversion of grazing commons and farmlands into urban development projects has been the most severe process in the northern and north-eastern parts of BNP (Jayaprakash & Hickey



**Figure 1.** Map indicating the location of Bannerghatta National Park.

2019). The villages in and around BNP face crop damage, livestock loss and even human injuries or losses due to wildlife despite the implementation of protective measures such as fencing and trenching.

### Data collection

Twelve villages located along the boundary of BNP were selected purposively to geographically cover the study area, representing different distances to the main cities and towns in the region. Within these villages, permanent residents were selected based on their age, caste and gender to ensure a balanced representation in the sample. Informants between 18 and 29 years old were considered young, those from 30 to 59 years of age were considered middle-aged and those aged 60 years and above were considered old. After stratification, informants were chosen by convenience sampling and approached at their homes and in key public locations, such as tea shops, *kattes* (sacred platforms that are popular meeting points), temples and marketplaces. The total number of respondents amounted to 489 (Appendix S1 shows the socio-demographic characteristics of the sample of respondents).

Face-to-face PPGIS surveys were conducted in May 2022 using *maptionnaire* software on tablets with the facilitation of five locally trained enumerators. The questionnaire (Appendix S2) was pre-tested in three different villages and subsequently adapted before starting the final survey. In a first questionnaire section, the respondents located their homes and those places in the landscape on the map that they associated with different uses and values related to BNP. The typology of uses and values (both here referred to as ‘values’) was built based on previous studies examining landscape values through PPGIS (Garcia-Martin et al. 2017). In the second section, respondents located specific places of disvalued (where people had experienced human-wildlife conflict, crop damage, risk to livestock and other nuisances). The final choice of categories was developed through a series of exploratory visits and

interviews with residents and visits to the area. The third section included two open questions that asked about the respondents’ most valued and most disvalued elements from their everyday landscapes. The last section of the questionnaire captured socio-demographic information. Initially, it took some time for older and illiterate respondents to familiarize themselves with the digital maps. However, the field researcher (PT) and enumerators assisted them in navigating the maps. Younger and more educated respondents were typically able to fill in the survey without assistance. Overall, we believe that our procedure allowed all respondents to accurately identify all of the locations on the maps. A systematic self-assessment provided by University of Kassel indicated that our survey should be performed using an ethically sensitive procedure (but the university’s ethics committee did not require formal ethical approval). All respondents were informed about the objectives of the study and their rights regarding data use. They consented explicitly before a survey was conducted.

### Data analysis

The answers to the free-listing questions regarding positively and negatively valued landscape elements were inductively analysed. The open responses were translated from the local language to English. For content analysis, a protocol was developed that included several iterations of coding. The items associated with the most valued elements related to BNP were grouped into eight categories, inspired by the well-being components of Rogers et al. (2012): ecosystems and biodiversity; social relationships; climate and ecosystem functions; work and leisure; health; education; agency and political voice; and physical and economic security. Items associated with the most disvalued elements were grouped into seven categories, partly following the typology of PA conflicts by Soliku & Schraml (2018): restricted access

**Table 1.** Relative proportions and frequencies of the most valued and most disvalued elements in the respondents' everyday landscapes. Relative proportions were calculated from the total of 1430 (for the most valued) and 908 (for the most disvalued) items. Only those individual elements coded at least 10 times are displayed.

Category	Element	n	
<b>Valued</b>			
Ecosystems and biodiversity (n = 804)	Agricultural land	239	
	Forest	193	
	Water bodies	121	
	Pleasant environment	79	
	Livestock	64	
	Village	39	
	Grazing grounds	14	
	Aesthetic beauty	10	
	Social relationships (n = 235)	Sense of place	191
		Community bonding	31
		Good neighbourhood	13
	Climate and ecosystem functions (n = 181)	Good air quality	85
		Pleasant weather	71
		Abundant rain	18
Work and leisure (n = 153)		Temple	87
	Katte	32	
	Good transportation and communication	22	
	Peacefulness and tranquillity	24	
Health (n = 26)			
Education (n = 14)			
Agency and political voice (n = 11)			
Physical and economic prosperity (n = 6)			
<b>Disvalued</b>			
Restricted area conflicts (n = 385)	Poor village infrastructures	145	
	Poor drainage and sanitation	91	
	Poor road access and communication	82	
	Poor electricity and water supply	50	
	Poor employment opportunities	10	
	Human-wildlife conflicts (n = 306)	Human-elephant conflicts	88
		Human-monkey conflicts	73
		Other human-wildlife conflicts	70
		Crop damage by wildlife	43
	Urbanization-conservation conflicts (n = 75)	Human-wild boar conflicts	20
Disturbance from mining		31	
Environmental degradation		28	
Agriculture and land-use conflicts (n = 72)		Restricted use and access to forest	40
	Land-use conflicts	32	
	Social coexistence conflicts (n = 51)	Waste dumping	18
Antisocial activities		33	
Indigenous rights conflicts (n = 12)		Lack of village commons	12
	Conflicts related to legislation and policy (n = 7)		

conflicts; human-wildlife conflicts; urbanization-conservation conflicts; agriculture and land-use conflicts; social coexistence conflicts; Indigenous rights conflicts; and conflicts related to legislation and policy.

The spatial extent and intensity of values and disvalues were analysed by calculating spatial indices (Fagerholm et al. 2021). Firstly, we created kernel density surfaces from the mapped point data separately for the uses and values and for the disvalues (Brown & Fagerholm 2015). Kernel density estimation was calculated as points per hectare, with a 200-m output cell size and 200-m search radius. Richness was calculated as the total number of values or disvalues present in each 200-m cell. Diversity was calculated with the Shannon diversity index ( $H'$ ), which measures the diversity and occurrence of values or disvalues in each 200-m cell; an  $H'$  value of zero indicates that only a single value or disvalue is present in the cell. The maximum values for  $H'$  are reached when all of the values or disvalues are represented by the same number of mapped points in a specific cell.

The identified values as well as disvalues were analysed using descriptive statistics. We then used Kruskal-Wallis tests to assess whether the values or disvalues identified varied according to the different socio-demographic characteristics of the respondents and between the different locations inside and outside BNP.

## Results

### Most valued and disvalued elements of respondents' everyday landscapes

The exploration of the most valued and disvalued elements resulted in a broad diversity of elements, which included 37 different most valued elements and 25 most disvalued elements (Appendices S3 & S4). Among the valued elements, the most frequently mentioned were agricultural land, forest, sense of place, water bodies, temples and good air quality (Table 1). When classifying these items based on the well-being dimension to which they referred, the most relevant category was ecosystems and biodiversity (Table 1). The most frequently mentioned disvalued elements were poor village infrastructure, poor sanitation systems, human-elephant conflicts, poor road access, human-monkey conflicts and other human-wildlife conflicts (Table 1). The most relevant categories were restricted area conflicts and human-wildlife conflicts (Table 1 & Appendix S4).

### Landscape values and disvalues mapped

The 489 respondents mapped a total 7846 values and 1949 disvalues (Table 2). On average, each respondent mapped 20

**Table 2.** Values and disvalues mapped inside and outside Bannerghatta National Park (BNP).

Value	Outside BNP		Inside BNP		Kruskal–Wallis test		
	No. of points	Mean $\pm$ SD	No. of points	Mean $\pm$ SD	Observed value	Critical value	P-value
Biodiversity	555	2.27 $\pm$ 1.63	532	2.17 $\pm$ 1.53	0.67	3.84	0.41
Soil fertility, air and water quality	492	2.02 $\pm$ 1.14	493	2.01 $\pm$ 1.15	0.10	3.84	0.76
Aesthetic value	474	1.94 $\pm$ 1.19	479	1.96 $\pm$ 1.23	0.01	3.84	0.97
Water provisioning	464	1.90 $\pm$ 0.90	469	1.91 $\pm$ 1.00	0.01	3.84	0.98
Spiritual values	429	1.76 $\pm$ 0.87	429	1.75 $\pm$ 0.83	0.01	3.84	0.98
Farming	366	1.50 $\pm$ 1.54	411	1.68 $\pm$ 1.63	2.44	3.84	0.12
Fodder	338	1.39 $\pm$ 0.94	309	1.26 $\pm$ 0.90	1.62	3.84	0.20
Social relations	281	1.15 $\pm$ 0.85	269	1.10 $\pm$ 0.78	0.39	3.84	0.53
Forest product harvesting	237	0.97 $\pm$ 0.95	235	0.96 $\pm$ 0.95	0.02	3.84	0.88
Fishing	202	0.83 $\pm$ 0.92	182	0.74 $\pm$ 0.94	1.88	3.84	0.17
Historical value	99	0.41 $\pm$ 0.55	101	0.41 $\pm$ 0.63	0.13	0.13	0.72
<i>Total values</i>	3937	17.19 $\pm$ 4.92	3909	17.19 $\pm$ 0.92	0.21	3.84	0.65
Crop damage by wildlife	387	1.59 $\pm$ 1.11	404	1.65 $\pm$ 1.06	0.52	3.84	0.47
Risk of wildfires	227	0.93 $\pm$ 1.96	199	0.81 $\pm$ 1.84	0.15	3.84	0.70
Personal harm by wildlife	171	0.70 $\pm$ 0.66	169	0.69 $\pm$ 0.70	0.14	3.84	0.71
Antisocial activities	102	0.42 $\pm$ 0.69	117	0.48 $\pm$ 0.80	0.23	3.84	0.63
Mining noise	92	0.38 $\pm$ 0.83	81	0.33 $\pm$ 0.76	0.67	3.84	0.41
<i>Total disvalues</i>	979	4.28 $\pm$ 3.18	970	4.28 $\pm$ 3.03	0.31	3.84	0.57

**Table 3.** Intensity, richness and diversity of values and disvalues as mapped inside and outside Bannerghatta National Park (BNP).

	Values		Disvalues	
	Inside BNP	Outside BNP	Inside BNP	Outside BNP
Area (ha)	4428	5076	2204	2664
<i>Intensity</i>				
Mean score $\pm$ SD	0.9 $\pm$ 2.2	0.7 $\pm$ 2.1	0.5 $\pm$ 0.8	0.9 $\pm$ 0.7
Range	0.000049–26.0	0.00000133–23.9	0.0000068–5.8	0.0000214–6.7
<i>Richness</i>				
Mean score $\pm$ SD	2.6 $\pm$ 2.1	2.5 $\pm$ 2.1	1.6 $\pm$ 0.9	1.4 $\pm$ 0.8
Range	1–11	1–12	1–5	1–6
<i>Diversity</i>				
Mean score $\pm$ SD	1.1 $\pm$ 0.4	1.2 $\pm$ 0.5	0.8 $\pm$ 0.2	0.7 $\pm$ 0.3
Range	0.4–2.2	0.5–2.4	0.2–1.4	0.2–1.7

locations. Biodiversity was the most frequently identified among the values, followed by regulation of diverse ecosystem functions (soil fertility, clean air and water), aesthetics and the supply of water. In relation to disvalues, crop damage inflicted by wildlife was the most frequently mapped disvalue, followed by risk posed by wildfire and personal damage inflicted by animals. No significant differences were found between areas inside and outside of BNP.

Overall, differences in intensity, richness and diversity of values and disvalues between those inside and outside BNP did not vary strongly; however, values and disvalues showed slightly higher mean richness values inside BNP compared to those outside. In contrast, the intensity of mapped disvalues was higher outside BNP (Table 3).

### Spatial pattern analysis

Values were generally spread over a larger area across the landscape (9504 ha) compared to disvalues (4868 ha) and showed predominantly higher mean intensity, richness and diversity scores of mapped sites (Table 3). Both values and disvalues were largely clustered within the villages and their surroundings.

Values and disvalues were spread across a smaller area inside BNP compared to outside. The values and disvalues of those outside BNP appeared to be slightly less clustered and were also mapped on locations near the National Park borders. While several

values were mapped in areas within BNP, hardly any values or disvalues were allocated to areas outside BNP beyond the immediate surroundings of respondents' villages.

Visual inspection showed a large spatial overlap among intensity, richness and diversity areas for both values and disvalues. However, high-intensity areas were larger in size than high-richness and high-diversity areas of values and disvalues.

### Relationship with socio-demographic information

Values and disvalues were mapped differently by different groups of respondents (Table 4). Men identified on average more locations than women. However, the only statistically significant gender differences were that men valued social activities in landscapes more strongly, while women mapped the disvalue of mining disturbances more frequently. Respondents self-identified as general caste perceived substantially more values than other respondents, while respondents from marginalized castes (so-called Other Backward Classes (OBC) and Scheduled Classes) perceived more disvalues. People working in agriculture generally mapped many more values and disvalues than people working in other sectors. Respondents who owned some land in the study area identified more values associated with the landscape, particularly certain key resources (farming land, water, fodder), than landless respondents. Lastly, respondents living closer to a city perceived more disvalues than respondents living far from cities.

**Table 4.** Kruskal–Wallis tests comparing the frequencies of mapped landscape values and disvalues based on socio-demographic information. Arrows indicate statistically significant differences ( $p < 0.05$ ). For instance, ↑ indicates that respondents working in agriculture mapped water-related values significantly more often than those who did not work in this sector.

	Gender		Age (years)			Caste					Work in agriculture		Ownership of land		Distance to city	
	Male	Female	18–29	30–59	≥60	General	OBC	Other	Scheduled Caste	Scheduled Tribe	Yes	No	Yes	No	1–15 km	>15km
Water	-	-	-	-	-	-	-	-	-	-	↑	↓	↑	↓	-	-
Farming	-	-	-	-	-	↑	↓	-	↓	-	↑	↓	↑	↓	-	-
Fodder	-	-	-	-	-	-	↑	-	↓	-	↑	↓	↑	↓	-	-
Forest product harvesting	-	-	-	-	-	-	-	-	-	-	↑	↓	-	-	-	-
Fishing	-	-	-	-	-	-	-	-	-	-	↑	↓	-	-	↓	↑
Social activities	↑	↓	-	-	-	↓	↑	-	-	-	-	-	-	-	-	-
Aesthetic value	-	-	-	-	-	↑	↓	-	-	-	↑	↓	-	-	-	-
Biodiversity	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-
Spiritual value	-	-	-	-	-	-	-	-	-	-	↑	↓	-	-	-	-
Historical value	-	-	-	-	-	-	-	-	-	-	-	-	↓	↑	-	-
Regulating services	-	-	-	-	-	↓	↑	↓	↓	↑	-	-	-	-	-	-
<i>Total values</i>	-	-	-	-	-	↑	-	-	-	-	↑	↓	-	-	-	-
Crop damages by animals	-	-	-	-	-	↑	↑	↓	↓	-	↑	↓	↑	↓	↑	↓
Personal damages by animals	-	-	-	-	-	↓	↑	↓	↓	↑	↑	↓	-	-	-	-
Mining disturbance	↓	↑	-	-	-	-	-	-	-	-	↑	↓	-	-	-	-
Antisocial activity	-	-	-	-	-	↑	↓	-	↑	-	↓	↑	-	-	-	-
Risk of wildfire	-	-	-	-	-	↑	↓	-	-	-	-	-	-	-	↑	↓
<i>Total disvalues</i>	-	-	-	-	-	-	↓	-	↓	-	↑	↓	-	-	↑	↓
<i>All</i>	↑	↓	-	-	-	-	-	-	-	-	↑	↓	-	-	-	-

OBC = 'Other Backward Classes'.

## Discussion

This study strives to elicit how residents living in close proximity to an urbanizing national park perceive landscape values and disvalues. In our survey, communities living in and those living around BNP mapped landscape values and disvalues rather similarly. However, we did identify differences in relation to the places where respondents mapped landscape values, which were smaller inside BNP, probably because more restrictions on movements and uses in this landscape apply. Correspondingly, we found a higher intensity of both values and disvalues inside BNP, possibly indicating a higher potential for conflict.

When our study asked about PA residents' subjective perceptions of those landscape features that supported their individual well-being, ecosystems and biodiversity (a category that includes many land-related items, including productive uses) emerged as the most relevant dimension, highlighting the importance of nature to local communities. Similar findings of people valuing predominantly biodiversity and ecosystem-related features were gained in a recent survey in Portuguese UNESCO biosphere reserves (Moreira et al. 2024). Second in importance was the stability of the ecosystems and their capacity to navigate disturbances, which relates to the importance of local farming and other natural resource uses for the communities' livelihoods. Similarly, communities living near nature reserves in China emphasized vegetation, clean water and air quality as central aspects of life satisfaction (Ma et al. 2021). Other dimensions of well-being featured very little in our study. In contrast, studies performed in Europe highlighted outdoor recreation and cultural values much more prominently (Fagerholm et al. 2020, Lampinen et al. 2024). People's positive and negative perceptions of how a PA influences their well-being are clearly mediated by local- and national-scale factors (Allendorf et al. 2019).

The most disvalued landscape features reflected the most pressing conflicts around BNP. They relate primarily to human-wildlife conflicts that – as in the case of crop raids by elephants – are exacerbated by urbanization pressures and their disruptive impacts on, for instance, wildlife corridors. These conflicts may be particularly significant around BNP due to its irregular shape, which results in an extended boundary area compared to the PA's core area. Such social impacts of PAs have strong effects on people's subjective well-being levels (Jones et al. 2020). However, the expressed negative aspects also went beyond factors directly influenced by the PA and included access to infrastructure and urbanization-related conflicts.

The fact that the most valued landscape features were linked to ecosystems and that the most disvalued features were related to PA management and spatial planning highlights the complex relationship of residents with their natural protected surroundings. Nature can be at the same time the most important factor contributing to well-being and the source of the most relevant problems, highlighting that the direction and magnitude of landscape valuation are location-dependent. For example, local residents appreciated values such as recreation and crop production in a watershed in the USA while at the same time being concerned about increased flooding, biological invasions and other social-ecological stressors of the watershed (van Riper et al. 2024). Similar findings were obtained in urban green spaces (Neidig et al. 2023) and forests (Baumeister et al. 2022). This type of paradox reflects the importance of considering the plural values of nature and the local context-related need for implementing methods that capture nature's both positive and negative impacts

on human well-being (De Vos et al. 2018). It also highlights the need to further consider horizontal interactions within the diverse values of nature – for example, how specific values vary not only between but also within the same subjects (Raymond et al. 2023).

The findings from our participatory mapping were similar to those from studies of perceived landscape values in PAs elsewhere in the world (e.g., Yee et al. 2021). The most frequently mapped landscape values were biodiversity, soil fertility, air and water quality and aesthetic values. Based on the comprehensive number of landscape values mapped, BNP appears – despite its focus on wildlife conservation – to be a multifunctional landscape that supports a broad range of uses and values simultaneously. A similarly multifunctional bundle centring around linking biodiversity and agricultural and heritage values was identified for a PA in Norway (Cusens et al. 2022). However, many disvalues were also mapped (although to a lesser degree), especially those related to human-wildlife conflicts. In addition to harm to crops, livestock and people, the prominence of wildfire risks was notable in BNP. This concern may reflect disagreement with current BNP management practices, which have led to increased shrubland and tree density, subsequently raising wildfire risks. The inclusion of such disvalues in PPGIS highlights critical areas for management intervention and community engagement to ensure that conservation efforts align more closely with the needs and perceptions of local residents (Baumeister et al. 2022).

The spatial overlap of values and disvalues that we found again suggests the importance of incorporating the mapping of negative perceptions in PPGIS to gain a more nuanced understanding of landscape dynamics. Although human-wildlife conflicts and debates over 'fortress conservation' are prevalent in the area, it is surprising that BNP-related landscape values were more frequently mapped than disvalues. This discrepancy might indicate a complex relationship between residents and BNP, where positive perceptions outweigh negative experiences, despite ongoing conflicts.

Similar to other PPGIS studies (Loc et al. 2021), our study revealed comparatively substantial differences between socio-demographic respondent groups. Caste, employment in agriculture and legal ownership of land were the most relevant factors that mediated human-nature relationships. These findings partly confirm and partly contrast with findings from previous socio-cultural valuation studies. In particular, individuals experiencing closer daily interaction with nature, such as those working in agriculture and those owning land, tended to associate more values and disvalues with the landscape, similar to what was found previously by Fagerholm et al. (2019). The fact that men observed more landscape values than women may be a consequence of the typical division of work within families, with men being more dedicated to outdoor work (most notably farming) and women to home-keeping (Thapa et al. 2023). We found several differences between more privileged and more marginalized castes that deserve more detailed exploration. Our results suggest that marginalized castes perceive disvalues from BNP more intensively. In other valuation studies in the area, general caste members typically expressed stronger appreciation for biodiversity and recreational values, whereas members of poorer and more marginalized castes focused more strongly on the provisioning of ecosystem services for subsistence (Nagendra 2016, Plieninger et al. 2022). The magnitude of these differences points to the importance of a differentiated analysis of local residents in and around PAs for conservation planning and management. Other intersecting respondent characteristics, such as age or level of urbanity, were less influential.



Our approach was mostly quantitative in nature and may not capture the more subtle or multifaceted landscape values (e.g., the spiritual meanings of nature) around the National Park. It would also be important to monitor changes in perceived values and disvalues over time with further ongoing urbanization. Expanding assessments of nature's values and disvalues by focusing more on the rights of nature, people's bonds to nature and people's trauma stemming from negative natural events and environmental degradation (Luque-Lora 2024) may be another avenue for future research.

## Conclusions

We highlight the need to develop specific strategies that can address the unique challenges facing PAs due to urbanization, especially in the Global South. Our participatory mapping study of urbanizing BNP (India) provides two main methodological contributions to the study of environmental conservation in this context. First, it expands participatory mapping to include both the values and disvalues of PAs. Our study indicates that local communities extensively use and value PAs for their well-being, but they also identify serious disvalues, including harm to humans. Such a consideration of both the positive and negative impacts of PAs on people is a prerequisite for balanced decision-making and, as a consequence, improved people-PA relationships. Second, this study integrates free-listing with participatory mapping, which is helpful for linking mapped values and disvalues with informants' relational connections to the landscape. By considering multiple values and methods, we can focus on a plurality of methods, actors and values in biodiversity conservation (Pascual et al. 2017). Our approach allowed us to uncover a more nuanced picture of landscape value perception, which would have been incomplete or biased if we had conducted this assessment using only one of these methods.

While the number of perceived values outweighs that of disvalues, many of the perceived landscape values are related to crop cultivation and pastoralism. Our study highlights that in and around PAs facing urbanization pressure, values can still be fundamentally tied to supporting basic food production needs and traditional livelihoods; this challenges assumptions regarding changing priorities in urbanizing landscapes. It is thus crucial to design PAs in a way that accommodates agricultural uses in their surroundings, especially within an urbanization context in which farming is compromised both by PA regulations and by the conversion of farmland to built-up land. At the same time, our study highlights that many of these disvalues can be addressed by improving access to basic infrastructure, especially for the more marginalized residents.

**Supplementary material.** To view supplementary material for this article, please visit <https://doi.org/10.1017/S0376892925100234>.

**Acknowledgements.** We are grateful to our enumerators and all residents who responded to our interview requests. We thank Samudyatha Ramanada and Sarang KT for supporting the field data collection, Katri Väänänen for assisting in PPGIS data analysis and Chaima Mobarak for the map of the study area.

**Financial support.** This research has received funding from the German Research Foundation (DFG) through grant number 421285339 (FOR2432/2). DB acknowledges support from the Department of Biotechnology (DBT) of the Government of India (grant number BT/IN/German/DFG/14/BVCR/2019).

**Competing interests.** The authors declare none.

**Ethical standards.** The authors assert that all procedures contributing to this work comply with applicable national and institutional ethical guidelines.

## References

- Adhikari S, Southworth J, Nagendra H (2015) Understanding forest loss and recovery: a spatiotemporal analysis of land change in and around Bannerghatta National Park, India. *Journal of Land Use Science* 10: 402–424.
- Allendorf TD, Radeloff VC, Keuler NS (2019) People's perceptions of protected areas across spatial scales. *Parks* 25: 25–38.
- Baumeister CF, Gerstenberg T, Plieninger T, Schraml U (2022) Geography of disservices in urban forests: public participation mapping for closing the loop. *Ecosystems and People* 18: 44–63.
- Beery T, Stahl Olafsson A, Gentin S, Maurer M, Stålhammar S, Albert C, et al. (2023) Disconnection from nature: expanding our understanding of human-nature relations. *People and Nature* 5: 470–488.
- Brill G, Anderson P, O'Farrell P (2017) Urban national parks in the Global South: linking management perceptions, policies and practices to water-related ecosystem services. *Ecosystem Services* 28: 185–195.
- Brown G, Fagerholm N (2015) Empirical PPGIS/PGIS mapping of ecosystem services: a review and evaluation. *Ecosystem Services* 13: 119–133.
- Cusens J, Barraclough AM, Måren IE (2022) Participatory mapping reveals biocultural and nature values in the shared landscape of a Nordic UNESCO Biosphere Reserve. *People and Nature* 4: 365–381.
- De Pourcq K, Thomas E, Elias M, Van Damme P (2019) Exploring park-people conflicts in Colombia through a social lens. *Environmental Conservation* 46: 103–110.
- De Vos A, Joana CB, Dirk R (2018) Relational values about nature in protected area research. *Current Opinion in Environmental Sustainability* 35: 89–99.
- Engen S, Runge C, Brown G, Fauchald P, Nilsen L, Hausner V (2018) Assessing local acceptance of protected area management using public participation GIS (PPGIS). *Journal for Nature Conservation* 43: 27–34.
- Fagerholm N, Martín-López B, Torralba M, Oteros-Rozas E, Lechner AM, Bieling C, et al. (2020) Perceived contributions of multifunctional landscapes to human well-being: evidence from 13 European sites. *People and Nature* 2: 217–234.
- Fagerholm N, Raymond CM, Olafsson AS, Brown G, Rinne T, Hasanzadeh K, et al. (2021) A methodological framework for analysis of participatory mapping data in research, planning, and management. *International Journal of Geographical Information Science* 35: 1848–1875.
- Fagerholm N, Torralba M, Moreno G, Girardello M, Herzog F, Aviron S, et al. (2019) Cross-site analysis of perceived ecosystem service benefits in multifunctional landscapes. *Global Environmental Change* 56: 134–147.
- García-Martin M, Fagerholm N, Bieling C, Gounaridis D, Kizos T, Printsmann A, et al. (2017) Participatory mapping of landscape values in a pan-European perspective. *Landscape Ecology* 32: 2133–2150.
- Gaston KJ, Jackson SF, Cantú-Salazar L, Cruz-Piñón G (2008) The ecological performance of protected areas. *Annual Review of Ecology, Evolution and Systematics* 39: 93–113.
- González-García A, Palomo I, Arboledas M, González JA, Múgica M, Mata R, Montes C (2022) Protected areas as a double edge sword: an analysis of factors driving urbanisation in their surroundings. *Global Environmental Change* 74: 102522.
- Gopalakrishna S, Somashekar R, Anand V, Varma S (2010) Asian elephant and Bannerghatta National Park in Eastern Ghats, southern India. *Gajah* 33: 47–52.
- Güneralp B, Seto KC (2013) Futures of global urban expansion: uncertainties and implications for biodiversity conservation. *Environmental Research Letters* 8: 014025.
- Gurney GG, Adams VM, Álvarez-Romero JG, Claudet J (2023) Area-based conservation: taking stock and looking ahead. *One Earth* 6: 98–104.
- Holmes G (2013) Exploring the relationship between local support and the success of protected areas. *Conservation and Society* 11: 72–82.
- Ioki K, Din NM, Ludwig R, James D, Hue SW, Johari SA, et al. (2019) Supporting forest conservation through community-based land use planning and participatory GIS – lessons from Crocker Range Park, Malaysian Borneo. *Journal for Nature Conservation* 52: 125740.

- Jayaprakash LG, Hickey GM (2019) Mistaking the map for the territory: what does the history of Bannerghatta National Park, India, tell us about the study of institutions? *Society & Natural Resources* 32: 1433–1450.
- Jones N, Malesios C, Kantartzis A, Dimitrakopoulos PG (2020) The role of location and social impacts of protected areas on subjective wellbeing. *Environmental Research Letters* 15: 114030.
- Joppa LN, Pfaff A (2009) High and far: biases in the location of protected areas. *PLoS ONE* 4: e8273.
- Lampinen J, García-Antúnez O, Jalkanen J, Havu M, Olafsson AS, Gulstrud NM, Raymond CM (2024) Social values for biodiversity and carbon flow in cities. *People and Nature* 6: 2374–2389.
- Lliso B, Lenzi D, Muraca B, Chan KMA, Pascual U (2022) Nature's disvalues: what are they and why do they matter? *Current Opinion in Environmental Sustainability* 56: 101173.
- Loc HH, Park E, Thu TN, Diep NTH, Can NT (2021) An enhanced analytical framework of participatory GIS for ecosystem services assessment applied to a Ramsar wetland site in the Vietnam Mekong Delta. *Ecosystem Services* 48: 101245.
- Loos J (2021) Reconciling conservation and development in protected areas of the Global South. *Basic and Applied Ecology* 54: 108–118.
- Luque-Lora R (2024) IPBES: three ways forward with frameworks of values. *Environmental Science & Policy* 159: 103827.
- Ma L, Qin Y, Zhang H, Zheng J, Hou Y, Wen Y (2021) Improving well-being of farmers using ecological awareness around protected areas: evidence from Qinling region, China. *International Journal of Environmental Research and Public Health* 18: 9792.
- McDonald RI, Mansur AV, Ascensão F, Colbert MI, Crossman K, Elmqvist T, et al. (2019) Research gaps in knowledge of the impact of urban growth on biodiversity. *Nature Sustainability* 3: 16–24.
- Moreira M, Frázão L, Cusens J, Castro P, Gouveia AC, da Silva AA, et al. (2024) Engaging participatory mapping for co-management and spatial depiction of nature's contributions to people in the Portuguese biosphere reserves. *Ecosystems and People* 20: 2422920.
- Nagendra H (2016) *Nature in the City: Bengaluru in the Past, Present, and Future*. Oxford, UK: Oxford University Press.
- Nagendra H, Bai X, Brondizio ES, Lwasa S (2018) The urban south and the predicament of global sustainability. *Nature Sustainability* 1: 341–349.
- Nagendra H, Sudhira HS, Katti M, Schewenius M (2013) Sub-regional assessment of India: effects of urbanization on land use, biodiversity and ecosystem services. In: T Elmqvist, M Fragkias, J Goodness, B Güneralp, PJ Marcotullio, RI McDonald, et al. (eds), *Urbanization, Biodiversity and Ecosystem Services: Challenges and Opportunities. A Global Assessment* (pp. 65–74). Berlin, Germany: Springer.
- Neidig J, Anguelovski I, Lliso B, Pascual U (2023) Pluralizing environmental values for urban planning: how to uncover the diversity of imaginaries about socio-natures from Vitoria-Gasteiz (Basque Country, Spain). *People and Nature* 5: 1262–1283.
- Oostvogels VJ, Dumont B, Nijland HJ, de Boer IJM, Ripoll-Bosch R (2024) What about the negatives? An integrated framework for revealing diverse values of nature and its conservation. *People and Nature* 6: 2633–2646.
- Palomo I, Martín-López B, Potschin M, Haines-Young R, Montes C (2013) National parks, buffer zones and surrounding lands: mapping ecosystem service flows. *Ecosystem Services* 4: 104–116.
- Palomo I, Montes C, Martín-López B, González JA, García-Llorente M, Alcorlo P, Mora MRG (2014) Incorporating the social-ecological approach in protected areas in the Anthropocene. *BioScience* 64: 181–191.
- Pascual U, Balvanera P, Díaz S, Pataki G, Roth E, Stenseke M, et al. (2017) Valuing nature's contributions to people: the IPBES approach. *Current Opinion in Environmental Sustainability* 26–27: 7–16.
- Plieninger T, Thapa P, Bhaskar D, Nagendra H, Torralba M, Zoderer BM (2022) Disentangling ecosystem services perceptions from blue infrastructure around a rapidly expanding megacity. *Landscape and Urban Planning* 222: 104399.
- Raymond CM, Anderson CB, Athayde S, Vatn A, Amin AM, Arias-Arévalo P, et al. (2023) An inclusive typology of values for navigating transformations towards a just and sustainable future. *Current Opinion in Environmental Sustainability* 64: 101301.
- Raymond CM, Cebrián-Piqueras MA, Andersson E, Andrade R, Schnell AA, Battioni Romanelli B, et al. (2022) Inclusive conservation and the Post-2020 Global Biodiversity Framework: tensions and prospects. *One Earth* 5: 252–264.
- Rogers DS, Duraiappah AK, Antons DC, Munoz P, Bai X, Fragkias M, Gutscher H (2012) A vision for human well-being: transition to social sustainability. *Current Opinion in Environmental Sustainability* 4: 61–73.
- Shackleton CM, Ruwanga S, Sinasson Sanni GK, Bennett S, De Lacy P, Modipa R, et al. (2016) Unpacking pandora's box: understanding and categorising ecosystem disservices for environmental management and human wellbeing. *Ecosystems* 19: 587–600.
- Solé L, Hearn KP, Witra T, Lechner AM, Fagerholm N (2025) Balancing landscape values and tourism choices: integrating participatory mapping and the IPBES Values Typology. *AMBIO* 54: 818–838.
- Soliku O, Schraml U (2018) Making sense of protected area conflicts and management approaches: a review of causes, contexts and conflict management strategies. *Biological Conservation* 222: 136–145.
- Thapa P, Torralba M, Bhaskar D, Nagendra H, Plieninger T (2023) Green in grey: ecosystem services and disservices perceptions from small-scale green infrastructure along a rural-urban gradient in Bengaluru, India. *Ecosystems and People* 19: 2223307.
- Torralba M, Topp E, Mounir F, Kmoch L, Plieninger T (2025) Diverging values and development options for agroforestry in a metropolitan area of North Africa. *Urban Forestry & Urban Greening* 104: 128633.
- van Riper CJ, Lorraine F, Ben L, Sanghyun L, Suresh S, Seunguk S, et al. (2024) Understanding the spatial dynamics of values and disvalues in the Kaskaskia River Watershed, USA through a social-ecological lens. *Ecosystems and People* 20: 2309391.
- Varma S, Anand VD, Gopalakrishna S, Avinash K, Nishant M (2009) Ecology, conservation and management of the Asian elephant in Bannerghatta National Park, southern India. *A Rocha India/ANCF: Asian Elephant Ecology and Conservation Reference Series* 1: 13–52.
- Venkataramana GV, Sreenivasa K, Lingaraju HG (2015) Evaluation of people's perceptions towards human-elephant conflict in and around Bannerghatta National Park. *Environment Conservation Journal* 16: 73–79.
- Wartmann FM, Acheson E, Purves RS (2018) Describing and comparing landscapes using tags, texts, and free lists: an interdisciplinary approach. *International Journal of Geographical Information Science* 32: 1572–1592.
- Yee JY, Loc HH, Poh YL, Vo-Thanh T, Park E (2021) Socio-geographical evaluation of ecosystem services in an ecotourism destination: PGIS application in Tram Chim National Park, Vietnam. *Journal of Environmental Management* 291: 112656.